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3 **Geography of global change and species richness in the North**

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16 **Abstract**

17 Different components of global change (e.g. climate change, land use, pollution and
18 introduced species) continue to alter biodiversity worldwide. As northern regions are still
19 relatively undisturbed and will likely face clear increases in temperature in the near-future,
20 we examined the signs of biodiversity change due to anthropogenic stressors using a
21 systematic review of previous studies. Our aim was to map where, in which way and due to
22 which stressor biodiversity in northern regions has changed. We made a systematic literature
23 search covering the years between 2000 and 2015 to obtain a comprehensive selection of
24 recent research. As species richness was clearly the most commonly used indicator of
25 biodiversity, we only concentrated on this aspect of biodiversity. We compared different
26 biological groups, regions and ecosystems. In the majority of the cases, anthropogenic
27 stressors had decreased species richness, or had no effects on it, while increasing or multiple
28 effects of stressors on species richness were less common. Freshwater ecosystems were most
29 sensitive to anthropogenic stressors, as species richness often decreased due to these
30 stressors. The effects of land use on richness were covered relatively widely in the selected
31 set of articles, but the effects of other components of global change on species richness
32 require further attention. Despite the fact that pollution was not as commonly studied stressor
33 as land use, it was the most harmful stressor type affecting species richness. Geographically,
34 most studies were located in boreal Canada or Fennoscandia, while no studies were executed
35 in vast circumpolar areas where the temperature rise has been greatest and the projected
36 climate change is likely to be fast. Overall, we could find an alarmingly small set of studies
37 that described the effects of actual anthropogenic stressors in real-life circumstances in
38 northern high latitudes.

39

40 **Keywords:** anthropogenic stress, Arctic, biodiversity, boreal, climate change, high latitudes

41 **Introduction**

42 The increase of human population size and massive consumption of natural resources have
43 led to the ongoing global environmental change. Components of human-induced global
44 change include, for instance, increased greenhouse gas concentrations in the atmosphere, land
45 use alteration, environmental pollution including nutrient loading, and introduction of non-
46 native species (Vitousek 1994; Chapin III et al. 2000). Some researchers have even proposed
47 that human activities have led to a new geological epoch (Table 1), the Anthropocene (Lewis
48 and Maslin 2015; Waters et al. 2016). During the Anthropocene, global changes have already
49 affected biodiversity in several and often intertwined ways (Chapin III et al. 2000; Butchart et
50 al. 2010; Vörösmarty et al. 2010; Maxwell et al. 2016). Moreover, numerous studies have
51 indicated that collapses of biodiversity due to global change may jeopardize ecosystem
52 functions and services essential to the life, societies and economies of humankind (Worm et
53 al. 2006; Cardinale et al. 2012; Vanbergen et al. 2013). Not even the generally sparsely-
54 populated northern high latitudes are safe from these human interventions and increased
55 utilisation of land and water, which form major disturbances to northern ecosystems (ACIA
56 2005; Halpern et al. 2008).

57 Emissions of greenhouse gases are currently highest in history. They have led to
58 changes in climate which have affected both human and natural systems as warming of the
59 atmosphere and oceans, reduction of snow and ice covers, and rising of the sea level. The
60 northern parts of the world will likely face the highest degrees of warming (IPCC 2013),
61 simultaneously affecting biodiversity in these vulnerable high-latitude regions (Chapin III et
62 al. 2000; Post et al. 2009). Climate change has already affected northern ecosystems and
63 biological communities in terrestrial (e.g. Aalto et al. 2014), freshwater (e.g. Nilsson et al.
64 2015) and marine realms (e.g. Kortsch et al. 2015). It has been predicted that by the end of
65 this century, climate change will be the most important driver of biodiversity change in the

66 arctic and boreal areas (Sala et al. 2000). The warming trend will likely change many aspects
67 of high-latitude biodiversity, such as distributions and abundances of species, the extent and
68 distribution of habitats, and introduction and spread of non-native species (e.g. ACIA 2005).

69 In addition to global warming, other human activities, such as timber harvesting,
70 agriculture and industry, have transformed northern high-latitude areas and changed the land
71 cover to a considerable degree (e.g. McGuire et al. 2007). Although there are still some
72 relatively pristine, large boreal forest areas left (e.g. Boonstra et al. 2016), boreal forests are
73 often shaped by silvicultural practices which modify the natural ecosystems of this vast
74 biome. The management of boreal forests alters the spatial qualities of these habitats, thus
75 possibly affecting occurrences of species that require continuous forest landscapes for
76 dispersal (Reunanen et al. 2010). Furthermore, development of industrial and urban
77 landscapes (e.g. ACIA 2005) and construction of roads (e.g. Trombulak and Frissell 2000)
78 can also cause habitat fragmentation, posing a severe threat to biodiversity (e.g. Hanski
79 2015). In the marine realm, the utilisation of marine areas for oil and gas drilling and fish
80 farming, for example, can change the ecosystem and biodiversity by altering habitat
81 conditions (e.g. ACIA 2005).

82 Species introductions to new areas by humans are closely linked to land use changes.
83 Hot-spots of alien species may occur near major human settlement areas (e.g. Wasowicz et al.
84 2013) or along roads (e.g. Trombulak and Frissell 2000). Alien species are typically
85 introduced to new regions unintentionally (e.g. Spaulding and Elwell 2007), but some species
86 are deliberately moved to new regions (e.g. Josefsson and Andersson 2001). In aquatic
87 ecosystems, major introduction pathways of alien species are the ballast waters of ships and
88 aquaculture (Molnar et al. 2008; Chan et al. 2014). Shipping-related transportation may
89 become a severe problem if ship traffic increases due to more pronounced melt of Arctic sea
90 ice (Stroeve et al. 2012; Eicken 2013). Overall, northern ecosystems are relatively species

91 poor, and the introduction of new species to these regions can severely affect native
92 biodiversity and ecosystem functioning in both terrestrial and aquatic realms (e.g. ACIA
93 2005). Importantly, the warming temperatures may favour alien species invasions while
94 disfavouring native species (e.g. Wasowicz et al. 2013).

95 Human activities also deteriorate air quality, which further affects biodiversity in high
96 latitudes. Airborne pollution and direct input of pollutants to ecosystems can alter both
97 terrestrial and aquatic biotas. Airborne carbon emissions threaten to acidify especially high-
98 latitude marine ecosystems (e.g. Steinacher et al. 2009). Pollution from industry, such as
99 mining and smelters, can also affect biodiversity at local scales (ACIA 2005; Zvereva and
100 Kozlov 2011). Nutrient enrichment can pose a threat to biodiversity especially in freshwater
101 (e.g. Vörösmarty et al. 2010) and marine ecosystems (e.g. ACIA 2005). Also, other types of
102 alien products entering the ecosystems exist. In marine ecosystems, pollution in the form of
103 plastic (e.g. Trevail et al. 2015), or fish feces and veterinary products from fish farms can be
104 released to the water. Furthermore, associated with shipping and oil industry, oil discharges
105 entering nature are relatively common (e.g. ACIA 2005).

106 Considering the variety of anthropogenic changes that Arctic and boreal areas have
107 already faced or will likely face in the near future, we systematically reviewed studies
108 conducted in these northern areas. Specifically, our aim was to map (1) where, (2) in which
109 way, and (3) due to which stressor biodiversity in northern regions has changed. We searched
110 for research papers that studied the effects of anthropogenic stressors on the diversity of
111 biological communities. We excluded all types of manipulative (including micro- and
112 mesocosm) studies in order to concentrate only on actual changes taking place at natural
113 spatial scales. To find out the main trends between the on-going global alterations and
114 biodiversity changes, we included all aquatic and terrestrial habitats, as well as all biological
115 groups in our systematic review. We will provide a general picture of where and what kind of

116 changes in biodiversity have already happened, or at least what has been studied by now. We
117 also address research gaps from spatial, organismal, ecosystem or stressor perspectives.

118

119 **Selection criteria and methods**

120 To find suitable articles, we selected appropriate keywords related to our themes of interest
121 and conducted a search in the Web of Knowledge (<http://apps.webofknowledge.com>). We
122 used three types of keywords: 1) words that describe the northern regions (arctic OR "high
123 latitude*" OR "high-latitude*" OR subarctic OR boreal OR polar); 2) words that are related
124 to global change (anthropogenic OR human* OR *pristine OR natural OR eutrophication OR
125 "nutrient enrichment" OR "habitat fragmentation" OR "land use*" OR "invasive species" OR
126 "alien species" OR acidification OR "climate change" OR "climate warming"), and 3) words
127 that are related to biodiversity (*diversit* OR richness OR evenness). We used these search
128 terms simultaneously and, for all rows, TOPIC was selected. We searched for articles
129 published between 2000 and 2015. The main search for suitable articles was done on
130 December 28, 2015 with a total of 3352 search results. To check if there were any more
131 articles added into the database matching with our search terms, one more search was done
132 on February 12, 2016 with a total of 3394 search results.

133 All authors were given an equal share of search results to go through and select articles
134 suitable for our scope. To ensure that the selection of articles was consistently made and to
135 guarantee high level of objectivity in the selection process, the first author double-checked all
136 selected articles. We selected articles that reported findings from northern areas (i.e. Arctic
137 and boreal regions) and dealt with the effects of anthropogenic stressors on biological
138 diversity of community-level data. In order to get selected, the article had to include a clear
139 comparative research layout (anthropogenic stress vs. no anthropogenic stress, anthropogenic
140 stress vs. natural stress, or an anthropogenic stress gradient). We attempted to include only

141 studies focusing on real-life situations, and thus we did not include any experimental or
142 manipulative studies. This is because we were interested to see whether there were any actual
143 trends reported regarding northern biodiversity change. We also did not include studies that
144 used a space-for-time substitution to illustrate the effects of e.g. climate change, studies that
145 tested ecological theories only, studies that did not have any clear stressors, purely predictive
146 studies, review papers or conference abstracts. These types of articles were numerous in the
147 initial search results and thus several exclusions were made. In addition, there were some
148 articles that did not clearly state their findings, and to refrain from making our own
149 deductions, we did not include such articles in the final set of articles either.

150 All authors collected information from articles that were likely suitable for comparative
151 purposes (see Table S1 in Supplementary material). Again, to ensure the uniform quality of
152 the data, the first author double-checked all collected information and made final decisions on
153 which articles to select. At this point, as it was clear that taxonomic richness was the most
154 commonly-used aspect of biodiversity in the selected papers, we decided to concentrate on
155 that aspect of biodiversity only. Richness was usually assessed at species level, so from now
156 on we use the term species richness to describe the taxonomic richness of the studies
157 included. The popularity of assessing species richness in the studies found is understandable
158 as species richness is the most commonly-measured aspect of biodiversity (Gaston 2000).

159 After the data were collected, we formed a number of categories from different
160 variables. For example, we formed five stressor type categories (i.e. climate change, land use,
161 pollution, introduced species and miscellaneous stressors; see Table S2). We also formed
162 nine major groups of biological organisms (i.e. plants, lichen, fungi, algae, bacteria,
163 invertebrates, fish, birds, mammals; see Table S3). The main terms we use along with
164 explanations are presented in Table 1. As we titled the five stressor type categories as
165 presented in Table 1, we did one more additional search for articles in the Web of Knowledge

166 (<http://apps.webofknowledge.com>) on July 20, 2016, to ensure that introduced species and
167 pollution were properly acknowledged in the search. The search terms were identical to the
168 original search apart from the second row, which had only two keywords on it (“introduced
169 species” OR pollution). There were 167 search results, which the first author checked and
170 selected articles if suitable for this review. Finally, the first author compiled a consistent final
171 data table including main information and variables from the final set of 90 selected articles
172 that fulfilled our criteria. As some of the selected articles studied species richness of multiple
173 biological groups, the final amount of separate data points in this review was 104. For the
174 final data table, see Table S4, and for the list of selected articles, see Table S5.

175 Our specific focus was to illustrate findings as cartographic presentations. For this
176 purpose, we used the continuous southern border of the boreal biome delineated using the
177 World Wildlife Fund terrestrial ecoregions map (Potapov et al. 2008) as the southern limit of
178 our research area. We also extrapolated this border to marine areas. To increase the amount
179 of cartographic information, we presented mean annual air temperature isotherms (Hijmans et
180 al. 2005) and NDVI (normalized difference vegetation index; Tucker 1979; Didan 2015) in
181 the maps as well. The approximate locations of the studies in the publications selected are
182 presented as a map in Fig. S6. The ID-number on the map and on the list of selected articles
183 (Table S5) is the connecting feature.

184

185 **Geographic clusters and gaps of research in the North**

186 We found 90 publications with 104 data points that passed our sieve. Most studies described
187 species richness-stressor relationships occurring in the southern provinces of Canada or
188 throughout Fennoscandia (Fig. 1, Fig. 2). In the continent of North America, vast Arctic
189 regions in Alaska and northern provinces territories of Canada have not been such thoroughly
190 studied in the context of species richness-global change relationships. Furthermore, our

191 systematic review showed that species richness-stressor relationships in Russia and the high
192 Arctic in general have been relatively seldom studied or they have been presented in non-
193 English and/or non-peer-reviewed publications. Thus, in that sense, almost the entire
194 circumpolar area presents a geographical research gap. As human activities, such as shipping,
195 oil extraction and mining, increase (AMAP 2012; Clement et al. 2013; Rhéaume and Caron-
196 Vuotari 2013), and as temperatures have been observed to rise in this area (IPCC 2013),
197 research needs to be focused on these still relatively natural, but constantly changing Arctic
198 areas. Importantly, the circumpolar research gap presents an area where climate warming is
199 predicted to be strongest compared to other parts of the world (IPCC 2013). Our map
200 illustrations (e.g. Fig. 1) show that there are few studies conducted in the region where the
201 mean annual air temperature is below -5°C (comparable to the zone of continuous permafrost
202 and extensive carbon pools; Schuur et al. 2015), thus representing a need for biodiversity
203 research focusing on especially cold environments. Likewise, most research has been focused
204 on the areas with high productivity indicated by NDVI in our maps. What is also important to
205 acknowledge when assessing species richness-stressor relationships in high-latitude regions is
206 the fact that these northern ecosystems go through four seasons, and biological organisms are
207 adapted to such change of seasons. Regarding climate change, especially winter temperatures
208 will likely increase the most, while summer temperatures are predicted to increase only
209 moderately (ACIA 2005; IPCC 2013). This seasonal difference in increasing temperatures
210 may further alter the complex relationships between components of global change and
211 biodiversity.

212

213 **Anthropogenic stress usually decreases or has no effects on species richness**

214 When considering the relationships between anthropogenic stressors and species richness of
215 different biological groups, negative effects of stressors on species richness were detectable

216 in one third of the cases (Fig. 1). Furthermore, one third of the cases showed no relationship
217 between species richness and any stressor, while increasing and multiple effects were clearly
218 less common. Anthropogenic global change thus affects species richness in various ways at
219 northern high latitudes, and not all effects are entirely negative or positive. This is
220 understandable as the relationships between biodiversity and stressors may be very complex
221 (e.g. Garcia et al. 2014), biotic interactions modify them (e.g. Schmitz et al. 2003; Olofsson
222 et al. 2013), biological communities may resist certain degrees of stress, or different stressors
223 have antagonistic effects on each other (Annala et al. 2014; Jackson et al. 2016). It is
224 however important to notice that increasing stress intensities or occurrences, probable in the
225 near future (ACIA 2005; Garcia et al. 2014; Nilsson et al. 2015), may affect species richness
226 in other, non-predictable ways. In addition, usually there are multiple stressors
227 simultaneously affecting biodiversity (ACIA 2005; Heino et al. 2009).

228 Geographically, there were some areas where species richness showed uniform
229 responses to human-induced stress (Fig. 1). For instance, in the Boreal Plains of western
230 Canada species richness usually had changed in some way due to anthropogenic stress. In
231 Fennoscandia, species richness seldom increased in response to human activities. Multiple
232 responses were more common in Fennoscandia than in North America. There were also some
233 areas (e.g. in south-eastern Canada) where species richness typically had not reacted to
234 anthropogenic stressors at all. In general, however, species richness throughout the northern
235 region showed several types of responses to different components of global change. In
236 addition, there were no clear trends observable between species richness responses and mean
237 annual air temperature or productivity. Further research conducted at the coldest latitudes or
238 areas with lower productivity might confirm or alter this finding.

239 **Terrestrial biodiversity most studied, but freshwater biodiversity most sensitive**

240 Terrestrial ecosystems were most commonly studied, with altogether 60 publications (Fig.
241 3a). We also found 27 publications on freshwater ecosystems, but only three publications on
242 marine ecosystems. Regarding the publications concentrating on terrestrial ecosystems, 70%
243 of the publications showed that species richness had changed due to human actions.
244 Terrestrial species richness had relatively evenly decreased, increased or exhibited multiple
245 responses due to anthropogenic stressors. Half of the studies conducted in freshwater
246 ecosystems, however, showed a negative relationship between species richness and an
247 anthropogenic stressor. Thus, it seems that freshwater biodiversity in northern regions is very
248 sensitive to different components of global change (see also Heino et al. 2009). Freshwater
249 species richness is, furthermore, more threatened in the future, as precipitation is predicted to
250 increase in the northern regions (IPCC 2013). The increasing rainfall may alter catchment
251 properties, ecosystem structure and function (ACIA 2005; Garssen et al. 2015; Lind et al.
252 2015).

253 Regarding marine ecosystems, all three studies showed that species richness had
254 changed due to human stress (Fig. 3a). Overall, we were surprised to find only few marine
255 studies dealing with anthropogenic effects on species richness. It is possible that such studies
256 do exist, but they were not captured with our search criteria or that those studies are simply
257 rare in northern regions. Moreover, marine systems differ remarkably from terrestrial and
258 freshwater systems, and thus traditional response-stressor studies may be more difficult to
259 conduct. Overall, the circumpolar research gap is at least partly linked to the absence of
260 marine studies. There is thus a need for studies focusing on marine species richness-stressor
261 relationships in northern high latitudes.

262 **Invertebrates and plants well covered in research**

263 For the entire northern region, most studies concentrated on species richness of either
264 invertebrates or plants (Fig. 1, Fig. 3b). Birds were also a relatively commonly-studied
265 biological group, followed by fungi and lichens which were more commonly studied in
266 Fennoscandia than in other northern areas. Species richness of fish and mammals were
267 surprisingly studied only in one paper each. Fish and mammals may be more commonly
268 studied as single species (Carey and Zimmerman 2014; Sonsthagen et al. 2014) and in
269 general ecological studies (Korsu et al. 2012; Hein et al. 2014), whereas studies on the effects
270 of stressors on their species richness seem to be less common at northern high latitudes.
271 Species richness of bacteria (i.e. richness of operational taxonomic units) was studied in two
272 publications only. Algae, containing traditionally-studied biological groups such as
273 phytoplankton, were neither also not studied very often in the context of anthropogenic
274 stressors and species richness. Perhaps nowadays algae are used for testing ecological
275 theories (e.g. Heino et al. 2010), or more complex indices than species richness are applied
276 (e.g. Lavoie et al. 2009).

277

278 **Different responses of species richness within and between biotic groups**

279 All biological groups that were studied more than once showed varying responses to
280 anthropogenic stressors (Fig. 1, Fig. 3b). In other words, the relationship between a stressor
281 and a biological group is not straightforward, but can be rather complex and probably
282 context-dependent (Sala et al. 2000; Woodward et al. 2010; Garcia et al. 2014). Again,
283 among many things, biotic interactions (e.g. Woodward, 2009), spatial scale (e.g. Garcia et
284 al. 2014) and regional characteristics (e.g. Bell et al. 2014) may affect the observed
285 relationships. For instance, in some study settings, although concentrating on one stressor
286 only while in fact multiple stressors were present, the effects of the stressor studied may be

287 attenuated (e.g. Ormerod et al. 2010). It is also possible that biological communities show
288 multiple responses to stress (Bell et al. 2014; Johnson and Angeler 2014).

289 The two most commonly-studied biological groups, plants and invertebrates, showed
290 somewhat different trends regarding species richness responses to anthropogenic stress.
291 Species richness of plants increased twice as often as species richness of invertebrates, which
292 in turn decreased twice as often compared to that of plants (Fig. 1, Fig. 3b). Additionally,
293 fungi and lichens, both present in terrestrial ecosystems, showed contrasting responses to
294 components of global change. Species richness of fungi more often showed decreasing
295 responses to anthropogenic stressors, whereas species richness of lichens usually did not react
296 to the stressors.

297

298 **Land use the most studied stressor, but pollution most harmful to species richness**

299 Land use, especially forestry, was the most studied stressor type over the entire northern
300 region (Fig. 2, Fig. 3c, S4). This is understandable because silviculture is a major human
301 activity across the vast boreal forest biome (e.g. Moen et al. 2014). Pollution was the second
302 most commonly-studied stressor type, followed by climate change and miscellaneous stressor
303 types. Climate change can be a particularly challenging stressor to study, because reliable
304 measurement of the effects typically requires a time-series of samples that is linked to
305 temperatures (see also Post et al. 2009). Miscellaneous stressor types included multiple
306 stressor types in our grouping. Importantly, as the situation with multiple stressors is
307 probably the most common in nature (ACIA 2005; Ormerod et al. 2010), there is a strong
308 need for studies that observe the effects of many simultaneously-acting stressors on species
309 richness (see also Post et al. 2009). From the major components of global change, introduced
310 species were the least studied stressor type in northern regions. More information is thus
311 needed on the effects of introduced species, as species introductions are predicted to increase

312 due to global change (e.g. Ware et al. 2014). Even though introduced species have surely
313 been studied, those studies typically concentrate on describing the distributional changes of
314 invasive species or pair-wise interactions between the introduced and some native species
315 (Leppäkoski and Olenin 2000; Hein et al. 2014).

316 Human-induced stressors can cause both positive and negative changes in biodiversity
317 (Garcia et al. 2014; Lind et al. 2014), which was also shown for northern areas in our
318 systematic review. For instance, land use showed approximately similar amounts of
319 increasing, multiple and decreasing effects on species richness, while the proportion of “no
320 effects” was pronounced when compared to the other stressor types. Climate change, in
321 general, showed multiple effects on species richness. Miscellaneous stressor types and
322 especially pollution usually decreased species richness (see also Zvereva and Kozlov 2011).
323 Consequently, the stressor types had different effects on species richness.

324

325 **Conclusions**

326 In northern regions, global change research on real-life species richness-stressor relationships
327 was surprisingly sparsely conducted both in quantity and in the spatial context. There were
328 vast areas where no research has been made, which is alarming as northern high latitudes will
329 likely face strongest changes due to global change (ACIA 2005; IPCC 2013). It is of course
330 possible that there were publications we could not find using the specific keywords, but we
331 are confident that the publications we included in this systematic review represent a good
332 selection of recent research conducted in northern ecosystems. Hence, we conclude that there
333 is a geographical research gap throughout the northern circumpolar area that deserves further
334 attention regarding the biodiversity-stressor relationships. Importantly, considering the
335 projected rate of future changes, the need for more research is urgent.

336 Overall, based on the publications reviewed, species richness had more commonly
337 changed due to an anthropogenic stressor than had remained unaffected by stressors. More
338 specifically, a decreasing trend of species richness was the most common type of response,
339 although there were also many types of other responses. Different biological groups showed
340 relatively similar distributions of responses in their species richness with a few exceptions. Of
341 the three different ecosystem types, species richness in freshwater ecosystems most often
342 showed a decrease in response to an anthropogenic stressor. This is an important finding for
343 policymakers to acknowledge. It is highly important to reduce the effects of stressors in these
344 ecosystems because the net effects are usually negative (e.g. Jackson et al. 2016).

345 Of the components of global change, land use change was clearly the most widely-
346 studied stressor type. Although not as commonly studied, pollution was most often related to
347 a decrease in species richness, thus posing a clear threat to species richness in northern high
348 latitudes. More research is needed on the species richness-stressor relationships regarding the
349 effects of climate change, introduced species and pollution. Surprisingly, studies addressing
350 the effects of multiple stressor types to biodiversity were exceptionally few. This trend
351 represents a need for more research focusing on multiple stressors acting in concert, which, in
352 the end, is the most common situation in nature (see also Halpern et al. 2008; Jackson et al.
353 2016; Titeux et al. 2016).

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359 **References**

- 360 Aalto, J., Venäläinen, A., Heikkinen, R.K., and Luoto, M. 2014. Potential for extreme loss in
361 high-latitude Earth surface processes due to climate change. *Geophys. Res. Lett.* **41**:
362 3914–3924.
- 363 ACIA. 2005. Arctic Climate Impact Assessment. Cambridge University Press, pp. 1042.
- 364 AMAP. 2012. Arctic Climate Issues 2011: Changes in Arctic Snow, Water, Ice and
365 Permafrost. Arctic Monitoring and Assessment Programme (AMAP), Oslo. SWIPA
366 2011 Overview Report, pp. 97.
- 367 Annala, M., Mykrä, H., Tolkkinen, M., Kauppila, T., and Muotka, T. 2014. Are biological
368 communities in naturally unproductive streams resistant to additional anthropogenic
369 stressors? *Ecol. Appl.* **24**: 1887–1897.
- 370 Bell, F.W., Hunt, S., Dacosta, J., Sharma, M., Larocque, G.R., Winters, J.A., and Newmaster,
371 S.G. 2014. Effects of Silviculture Intensity on Plant Diversity Response Patterns in
372 Young Managed Northern Temperate and Boreal Forests. *Ecoscience* **21**: 327–339.
- 373 Boonstra, R., Andreassen, H.P., Boutin, S., Hušek, J., Ims, R.A., Krebs, C.J., Skarpe, C., and
374 Wabakken, P. 2016. Why Do the Boreal Forest Ecosystems of Northwestern Europe
375 Differ from Those of Western North America? *BioScience*. doi:
376 10.1093/biosci/biw080.
- 377 Butchart, S.H.M., Walpole, M., Collen, B., van Strien, A., Scharlemann, J.P., Almond, R.E.,
378 Baillie, J.E., Bomhard, B., Brown, C., Bruno, J., Carpenter, K.E., Carr, G.M., Chanson,
379 J., Chenery, A.M., Csirke, J., Davidson, N.C., Dentener, F., Foster, M., Galli, A.,
380 Galloway, J.N., Genovesi, P., Gregory, R.D., Hockings, M., Kapos, V., Lamarque, J.F.,
381 Leverington, F., Loh, J., McGeoch, M.A., McRae, L., Minasyan, A., Hernández
382 Morcillo, M., Oldfield, T.E., Pauly, D., Quader, S., Revenga, C., Sauer, J.R., Skolnik,
383 B., Spear, D., Stanwell-Smith, D., Stuart, S.N., Symes, A., Tierney, M., Tyrrell, T.D.,

384 Vié, J.C., and Watson, R. 2010. Global biodiversity: indicators of recent declines.
385 *Science* **328**: 1164–1168.

386 Cardinale, B.J., Duffy, E., Gonzalez, A., Hooper, D.U., Perrings, C., Venail, P., Narwani, A.,
387 Mace, G.M., Tilman, D., Wardle, D.A., Kinzig, A.P., Daily, G.C., Loreau, M., Grace,
388 J.B., Larigauderie, A., Srivastava, D., and Naeem, S. 2012 Biodiversity loss and its
389 impacts on humanity. *Nature* **486**: 59–67.

390 Carey, M.P., and Zimmerman, C.E. 2014. Physiological and ecological effects of increasing
391 temperature on fish production in lakes of Arctic Alaska. *Ecol. Evol.* **4**: 1981–1993.

392 Chapin III, F.S., Zavaleta, E.S., Eviner, V.T., Naylor, R.L., Vitousek, P.M., Reynolds, H.L.,
393 Hooper, D.U., Lavorel, S., Sala, O.E., Hobbie, S.E., Mack, M.C., and Díaz, S. 2000.
394 Consequences of changing biodiversity. *Nature* **405**: 234–242.

395 Chan, F.T., Briski, E., Bailey, S.A., and MacIsaac, H.J. 2014. Richness–abundance
396 relationships for zooplankton in ballast water: temperate versus Arctic comparisons.
397 *ICES Journal of Marine Science*. doi: 10.1093/icesjms/fsu020.

398 Clement, J.P., Bengtson, J.L., and Kelly, B.P. 2013. Managing for the future in a rapidly
399 changing Arctic. Interagency Working Group on Coordination of Domestic Energy
400 Development and Permitting in Alaska.

401 Didan, K. 2015. MOD13C1 MODIS/Terra Vegetation Indices 16-Day L3 Global 0.05Deg
402 CMG V006. NASA EOSDIS Land Processes DAAC. doi:
403 10.5067/MODIS/MOD13C1.006.

404 Eicken, H. 2013. Arctic sea ice needs better forecasts. *Nature* **497**: 431–433.

405 Garcia, R.A., Cabeza, M., Rahbek, C., and Araujo, M.B. 2014. Multiple dimensions of
406 climate change and their implications for biodiversity. *Science* **344**: 1247579.

407 Garssen, A.G., Baattrup-Pedersen, A., Voesenek, L.A., Verhoeven, J.T., and Soons, M.B.
408 2015. Riparian plant community responses to increased flooding: a meta-analysis.
409 *Glob. Chang. Biol.* **21**: 2881–2890.

410 Gaston, K.J. 2000. Global patterns in biodiversity. *Nature* **405**: 220–227.

411 Halpern, B.S., Walbridge, S., Selkoe, K.A., Kappel, C.V., Micheli, F., D'Agrosa, C., Bruno,
412 J.F., Casey, K.S., Ebert, C., Fox, H.E., Fujita, R., Heinemann, D., Lenihan, H.S.,
413 Madin, E.M., Perry, M.T., Selig, E.R., Spalding, M., Steneck, R., and Watson, R. 2008.
414 A global map of human impact of marine ecosystems. *Science* **319**: 948–952.

415 Hanski, I. 2015. Habitat fragmentation and species richness. *J. Biogeogr.* **42**: 989–994.

416 Hein, C.L., Öhlund, G., and Englund, G. 2014. Fish introductions reveal the temperature
417 dependence of species interactions. *Proc. R. Soc. B* **281**: 20132641. doi:
418 10.1098/rspb.2013.2641

419 Heino, J., Bini, L.M., Karjalainen, S.M., Mykrä, H., Soininen, J., Vieira, L.C.G., and Diniz-
420 Filho, J.A.F. 2010. Geographical patterns of micro-organismal community structure:
421 are diatoms ubiquitously distributed across boreal streams? *Oikos* **119**: 129–137.

422 Heino, J., Virkkala, R., and Toivonen, H. 2009. Climate change and freshwater biodiversity:
423 detected patterns, future trends and adaptations in northern regions. *Biol. Rev.* **84**: 39–
424 54.

425 Hijmans, R.J., Cameron, S.E., Parra, J.L., Jones, P.G., and Jarvis, A. 2005. Very high
426 resolution interpolated climate surfaces for global land areas. *Int. J. Climatol.* **25**: 1965–
427 1978.

428 IPCC. 2013. Climate Change 2013. World Meteorological Organization. Fifth Assessment
429 Report.

430 Jackson, M.C., Loewen, C.J.G., Vinebrooke, R.D., and Chimimba, C.T. 2016. Net effects of
431 multiple stressors in freshwater ecosystems: a meta-analysis. *Glob. Chang. Biol.* **22**:
432 180–189.

433 Johnson, R.K., and Angeler, D.G. 2014. Effects of agricultural land use on stream
434 assemblages: Taxon-specific responses of alpha and beta diversity. *Ecol. Indic.* **45**:
435 386–393.

436 Joseffson, M., and Andersson, B. 2001. The Environmental Consequences of Alien Species
437 in the Swedish Lakes Mälaren, Hjälmaren, Vänern and Vättern. *AMBIO* **30**: 514–521.

438 Korsu, K., Heino, J., Huusko, A., and Muotka, T. 2012. Specific niche characteristics
439 facilitate the invasion of an alien fish invader in boreal streams. *International Journal of*
440 *Ecology*. doi:10.1155/2012/813016.

441 Kortsch, S., Primicerio, R., Fossheim, M., Dolgov, A.V., and Aschan, M. 2015. Climate
442 change alters the structure of arctic marine food webs due to poleward shifts of boreal
443 generalists. *Proc. R. Soc. B* **282**: 20151546.

444 Lavoie, I., Hamilton, P.B., Wang, Y-K., Dillon, P.J., and Campeau, S. 2009. A comparison of
445 stream bioassessment in Québec Canada using six European and North American
446 diatom-based indices. *Nova Hedwigia* **135**: 37–56.

447 Leppäkoski, E., and Olenin, S. 2000. Non-native Species and Rates of Spread: Lessons from
448 the Brackish Baltic Sea. *Biol. Invasions* **2**: 151–163.

449 Lewis, S.L., and Maslin, M.A. 2015. Defining the Anthropocene. *Nature* **519**: 171–180.

450 Lind, L., Nilsson, C., Polvi, L.E., and Weber, C. 2014. The role of ice dynamics in shaping
451 vegetation in flowing waters. *Biol. Rev.* **89**: 791–804.

452 Maxwell, S.L., Fuller, R.A., Brooks, T.M., and Watson, J.E.M. 2016. Biodiversity: The
453 ravages of guns, nets and bulldozers. *Nature* **536**: 143–145.

454 McGuire, A.D., Chapin III, F.S., Wirth, C., Apps, M.J., Bhatti, J.S., Callaghan, T.,
455 Christensen, T.R., Clein, J.S., Fukuda, M., Maximov, T., Omuchin, A., Shvidenko, A.,
456 and Vaganov, E. 2007. Responses of high latitude ecosystems to global change:
457 potential consequences for the climate system. *In Terrestrial Ecosystems in a Changing*
458 *World. Edited by Canadell, J.G., Pataki, D.E., and Pitelka, L.F. The IGBP Series,*
459 *Springer, pp. 297–310.*

460 Moen, J., Rist, L., Bishop, K., Chapin, F.S., Ellison, D., Kuuluvainen, T., Petersson, H.,
461 Puettmann, K.J., Rayner, J., Warkentin, I.G., and Bradshaw, C.J.A. 2014. Eye on the
462 Taiga: Removing global policy impediments to safeguard the boreal forest. *Conserv.*
463 *Lett.* **7**: 408–418.

464 Molnar, J.L., Gamboa, R.L., Revenga, C., and Spalding, M.D. 2008. Assessing the global
465 threat of invasive species to marine biodiversity. *Front. Ecol. Environ.* **6**: 485–492.

466 Nilsson, C., Polvi, L.E., and Lind, L. 2015. Extreme events in streams and rivers in arctic and
467 subarctic regions in an uncertain future. *Freshwater Biol.* **60**: 2535–2546.

468 Olofsson, J., te Beest, M., and Ericson, L. 2013. Complex biotic interactions drive long-term
469 vegetation dynamics in a subarctic ecosystem. *Philos. Trans. R. Soc. B.* doi:
470 10.1098/rstb.2012.0486.

471 Ormerod, S.J., Dobson, M., Hildrew, A.G., and Townsend, C.R. 2010. Multiple stressors in
472 freshwater ecosystems. *Freshwater Biol.* **55**: 1–4.

473 Post, E., Forchhammer, M.C., Bret-Harte, M.S. Callaghan, T.V., Christensen, T.R., Elberling,
474 B., Fox, A.D., Gilg, O., Hik, D.S., Høye, T.T., Ims, R.A., Jeppesen, E., Klein, D.R.,
475 Madsen, J., McGuire, A.D., Rysgaard, S., Schindler, D.E., Stirling, I., Tamstorf, M.P.,
476 Tyler, N.J., van der Wal, R., Welker, J., Wookey, P.A., Schmidt, N.M., and Aastrup, P.
477 2009. Ecological dynamics across the Arctic associated with recent climate change.
478 *Science* **325**: 1355–1358.

479 Potapov, P., Hansen, M.C., Stehman, S.V., Loveland, T.R., and Pittman, K. 2008. Combining
480 MODIS and Landsat imagery to estimate and map boreal forest cover loss. *Remote*
481 *Sens. Environ.* **112**: 3708–3719.

482 Reunanen, P., Fall, A., and Nikula, A. 2010. Biodiversity and ecological forest-cover
483 domains in boreal landscapes. *Biodivers. Conserv.* **19**: 665–678.

484 Rhéaume, G., and Caron-Vuotari, M. 2013. The future of mining in Canada's north. The
485 Conference Board of Canada.

486 Sala, O.E, Chapin III, F.S., Armesto, J.J., Berlow, E., Bloomfield, J., Dirzo, R., Huber-
487 Sanwald, E., Huenneke, L.F., Jackson, R.B., Kinzig, A., Leemans, R., Lodge, D.M.,
488 Mooney, H.A., Oesterheld, M., Poff, N.L., Sykes, M.T., Walker, B.H., Walker, M., and
489 Wall, D.H. 2000. Global biodiversity scenarios for the year 2100. *Science* **287**: 1770–
490 1774.

491 Schmitz, O.J., Post, E., Burns, C.E., and Johnston, K.M. 2003. Ecosystem responses to global
492 climate change: Moving beyond color mapping. *BioScience* **53**: 1199–1205.

493 Schuur, E.A.G., McGuire, A.D., Schädel, C. Grosse, G., Harden, J.W., Hayes, D.J., Hugelius,
494 G., Koven, C.D., Kuhry, P., Lawrence, D.M., Natali, S.M., Olefeldt, D., Romanovsky,
495 V.E., Schaefer, K., Turetsky, M.R., Treat, C.C., and Vonk, J.E. 2015. Climate change
496 and the permafrost carbon feedback. *Nature* **520**: 171-179.

497 Sonsthagen, S.A., Fales, K., Jay, C.V., Sage, G.K., and Talbot, S.L. 2014. Spatial variation
498 and low diversity in the major histocompatibility complex in walrus (*Odobenus*
499 *rosmarus*). *Polar Biol.* **37**: 497–506.

500 Spaulding, S., and Elwell, L. 2007. Increase in nuisance blooms and geographic expansion of
501 the freshwater diatom *Didymosphenia geminata*: Recommendations for response. US
502 Environmental Protection Agency.

503 Steinacher, M., Joos, F., Frölicher, T.L., Plattner, G-K., and Doney, S.C 2009. Imminent
504 ocean acidification in the Arctic projected with the NCAR global coupled carbon cycle-
505 climate model. *Biogeosciences* **6**: 515–533.

506 Stroeve, J.C., Serreze, M.C., Holland, M.M., Kay, J.E., Malanik, J., and Barrett, A.P. 2012.
507 The Arctic’s rapidly shrinking sea ice cover: a research synthesis. *Clim. Chang.* **110**:
508 1005–1027.

509 Titeux, N., Henle, K., Mihoub, J-P., and Brotons, L. 2016. Climate change distracts us from
510 other threats to biodiversity. *Front. Ecol. Environ.* **14**: 291.

511 Trevail, A.C., Gabrielsen, G.W., Kühn, S., and Van Franeker, J.A. 2015. Elevated levels of
512 ingested plastic in a high Arctic seabird, the northern fulmar (*Fulmarus glacialis*). *Polar*
513 *Biol.* **38**: 975–981.

514 Trombulak, S.C., and Frissell, C.A. 2000. Review of ecological effects of roads on terrestrial
515 and aquatic communities. *Conserv. Biol.* **14**: 18–30.

516 Tucker, C.J. 1979. Red and photographic infrared linear combinations for monitoring
517 vegetation. *Remote Sens. Environ.* **8**: 127–150.

518 Vanbergen, A.J., and the Insect Pollinators Initiative. 2013. Threats to an ecosystem service:
519 pressures on pollinators. *Front. Ecol. Environ.* **11**: 251–259.

520 Vitousek, P.M. 1994. Beyond global warming: ecology and global change. *Ecology* **75**:
521 1861–1876.

522 Vörösmarty, C.J., McIntyre, P.B., Gessner, M.O., Dudgeon, D., Prusevich, A., Green, P.,
523 Glidden, S., Bunn, S.E., Sullivan, C.A., Liermann, C.R., and Davies, P.M. 2010. Global
524 threats to human water security and river biodiversity. *Nature* **467**: 555–561.

525 Ware, C., Berge, J., Sundet, J.H., Kirkpatrick, J.B., Coutts, A.D.M., Jelmert, A., Olsen, S.M.,
526 Floerl, O., Wisz, M.S., and Alsos, I.G. 2014. Climate change, non-indigenous species

527 and shipping: assessing the risk of species introduction to a high-Arctic archipelago.
528 *Divers. Distrib.* **20**: 10–19.

529 Wasowicz, P., Przedpelska-Wasowicz, E.M., and Kristinsson, H. 2013. Alien vascular plants
530 in Iceland: Diversity, spatial patterns, temporal trends, and the impact of climate
531 change. *FLORA* **208**: 648–673.

532 Waters, C.N., Zalasiewicz, J., Summerhayes, C., Barnosky, A.D., Poirier, C., Galuszka, A.,
533 Cearreta, A., Edgeworth, M., Ellis, E.C., Ellis, M., Jeandel, C., Leinfelder, R., McNeill,
534 J.R., Richter, D.deB., Steffen, W., Syvitski, J., Vidas, D., Waple, M., Williams, M.,
535 Zhisheng, A., Grinewald, J., Odada, E., Oreskes, N., and Wolfe, A.P. 2016. The
536 Anthropocene is functionally and stratigraphically distinct from the Holocene. *Science*
537 **351**: 10.1126/science.aad2622.

538 Woodward, G. 2009. Biodiversity, ecosystem functioning and food webs in fresh waters:
539 assembling the jigsaw puzzle. *Freshwater Biol.* **54**: 2171–2187.

540 Woodward, G., Perkins, D.M., and Brown, L.E. 2010. Climate change and freshwater
541 ecosystems: impacts across multiple levels of organization. *Philos. Trans. R. Soc. B*
542 **365**: 2093–2106.

543 Worm, B., Barbier, E.B., Beaumont, N., Duffy, J.E., Folke, C., Halpern, B.S., Jackson, J.B.,
544 Lotze, H.K., Micheli, F., Palumbi, S.R., Sala, E., Selkoe, K.A., Stachowicz, J.J., and
545 Watson, R. 2006. Impacts of biodiversity loss on ocean ecosystem services. *Science*
546 **314**: 787–790.

547 Zvereva, E.L., and Kozlov, M.V. 2011. Impacts of Industrial Polluters on Bryophytes: a
548 Meta-analysis of Observational Studies. *Water Air Soil Poll.* **218**: 573–586.

549 Table 1. A glossary of the main concepts used in this systematic review.

Concept	Description
Northern region	Areas north from the southern border of the continuous boreal biome (delineated according to Potapov et al. 2008).
Species richness	We refer to all measures of taxonomic richness as species richness, because species level was the most studied taxonomic level in the publications.
Anthropocene	Our current epoch, which witnesses the overarching impacts of anthropogenic stressors on our planet's geology and ecosystems (Waters et al. 2016).
Global change	All anthropogenic actions that have led to a global change of the Earth.
Climate change	Human-induced climate warming.
Land use change	All kinds of anthropogenic landscape alterations (e.g. forestry, road building).
Pollution	Any non-natural matter that enters natural ecosystems due to human actions (e.g. nutrients, noise, road salt).
Introduced species	Alien species introduced to a new area due to human actions.
Miscellaneous stressors	Miscellaneous stressors (e.g. water regulation, recreation, wildlife management) or multiple stressor types.

550 **Figure captions**

551 **Fig. 1.** A map illustrating where and in which way species richness of different biological
552 groups has responded to components of global change. The thick grey line indicates the
553 northern region with the southern limit determined by the extent of the continuous boreal
554 zone (Potapov et al. 2008), which is also extrapolated to marine areas. Mean annual air
555 temperature isotherms are presented as solid lines in the map (red line: +5°C, purple line 0°C,
556 blue line -5°C; Hijmans et al. 2005). The background color of the map indicates productivity:
557 light green indicates high values of the normalized difference vegetation index (NDVI) and
558 light orange indicates low NDVI (there is no information on NDVI available from white
559 areas; Didan 2015).

560 **Fig. 2.** A map illustrating where and in which way different components of global change
561 have affected species richness. The thick grey line indicates the northern region with the
562 southern limit determined by the extent of the continuous boreal zone (Potapov et al. 2008),
563 which is also extrapolated to marine areas. Mean annual air temperature isotherms are
564 presented as solid lines in the map (red line: +5°C, purple line 0°C, blue line -5°C; Hijmans
565 et al. 2005). The background color of the map indicates productivity: light green indicates
566 high values of the normalized difference vegetation index (NDVI) and light orange indicates
567 low NDVI (there is no information on NDVI available from white areas; Didan 2015).

568 **Fig. 3.** A general picture of how much and which ecosystems (a), biological groups (b) and
569 stressor types (c) have been studied in the context of the species richness-anthropogenic
570 stress relationship, and how species richness has changed due to anthropogenic stressors.

571 **Supplementary material**

572 S1. Preliminary variables collected from articles.

573 S2. Main stressor categories and what they include.

574 S3. Biological groups and what they include.

575 S4. Final information collected from the selected articles.

576 S5. The list of selected articles.

577 S6. A map presenting the approximate locations of the studies reviewed.