

1 **Aminomethylation of spruce tannins and their application as coagulants for**  
2 **water clarification**

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11

12 **Abstract**

13 This study explored the potential for using tannin extract from spruce bark (*Picea abies*) as a  
14 coagulant. Spruce tannins were extracted with hot water and pulverized through spray drying and  
15 freeze drying. The pulverized tannins were cationized via the Mannich reaction with formaldehyde  
16 and diethanolamine or ethanolamine, and the coagulation performances of the spruce tannin  
17 coagulants were compared with an industrially extracted quebracho tannin, which was cationized  
18 with the same method. Jar test experiments with kaolin/river water indicated that all of the tannin  
19 coagulants were able to enhance particle settling significantly, although the quebracho tannin  
20 coagulants were slightly more efficient than the spruce tannin coagulants. Since the phenolic groups  
21 play the major role in the modification, the higher coagulation efficiency was probably related to the  
22 higher amount of proanthocyanidin in the quebracho tannin, and subsequently the higher cationic  
23 charge density obtained for the quebracho tannin coagulants. Nevertheless, the spruce tannin  
24 coagulants could still be considered to be effective products because they possessed more stable

25 turbidity and a total surface charge close to zero over a wide dosage range. The ethanolamine was a  
26 better amine source, at least for the coagulation of river water with a positive charge demand.  
27 Moreover, the study established the importance of charge density as an essential indicator for  
28 coagulant performance.

29

30 **Keywords:** Spruce bark tannin extract, Bio-coagulants, Mannich reaction, Charge density, Spray  
31 drying, Freeze drying.

32

### 33 **1. Introduction**

34

35 One of the most prominent environmental footprints of rapid industrialization is the production of a  
36 large volume of industrial effluent. These effluents contain a high concentration of suspended solids  
37 and organic materials which need to be eliminated before they can be discharged into the  
38 environment. One of the essential processes during industrial wastewater treatment is coagulation-  
39 flocculation, owing to its efficiency in reducing turbidity, dissolved organic matter and colour [1, 2,  
40 3]. It has been suggested that these aesthetic qualities are improved through two primary removal  
41 mechanisms: namely, charge neutralization and sweep coagulation. In charge neutralization, the  
42 positive charges from a positively charged coagulant are attracted to the negative charges of the  
43 colloids by electrostatic interaction. Precipitable flocs are then formed due to particle collision and  
44 agglomeration, induced by electrostatic charge neutralization. During sweep coagulation, organic  
45 matter/colloidal contaminants are entrapped on the surface of the insoluble hydroxide metal or  
46 organic polymer and are swept down with the precipitates as the suspension settles [4]. Besides charge  
47 neutralization and sweep coagulation, patch coagulation and bridging flocculation can also be utilized  
48 in the coagulation-flocculation process [5]. Patch coagulation mechanism refers to a mechanism in

49 which a polymer adsorbs on the particle surface and forms cationic patches. Subsequently, the  
50 oppositely charged patches of the particle are attracted to each other. When the polymer forms loops  
51 and tails, which are able to attach to other particles, the mechanism is called bridging flocculation.  
52 The effectiveness in improving aesthetic water quality by conventional chemical-based coagulants is  
53 well documented in the literature. However, the applicability of these coagulants has been marred by  
54 some disadvantages which include high chemical cost, generation of a large volume of non-  
55 degradable sludge that requires further processing thus increasing operational cost, production of a  
56 toxic by-product with detrimental effects to human health and the environment [6]. Other setbacks  
57 are ineffectiveness in cold waters and an increase in the acidity of treated waters [7,8]. Metal  
58 coagulants are often effective over a limited pH range, partly due to the optimal region for the  
59 formation of the charged species [9]. Therefore, it is imperative that efforts are made to develop more  
60 eco-friendly coagulants that could replace these conventional coagulants.

61 Bio-based coagulants provide a viable alternative to chemical-based coagulants because they possess  
62 the ability to counteract the afore-mentioned concerns. Bio-coagulants are highly degradable and  
63 produce a non-toxic and lesser volume of sludge [10]. Another advantage of bio-based coagulants  
64 over chemical-based is that raw materials can be sourced locally. For example, the use of indigenous  
65 plants will help eliminate the cost of purchasing and importing chemicals, especially in developing  
66 countries [11]. One type of bio-coagulant that has attracted more commercial interest in recent years  
67 is tannin-based polymers. This has been mainly attributed to their low cost, as tannins are widely  
68 distributed in the bark of many tree species, and to their effectiveness over a wide range of pH. For  
69 example, Grenda et al. [12] showed that tannin-based coagulant recorded impressive color removal  
70 levels for some synthetic dye water samples over a pH range of 1-14, a feat that has been reported as  
71 impossible with most metal salts. They are also worthy of consideration owing to their simple  
72 modification procedures, which include polymerization and aminomethylation. The  
73 aminomethylation process is known as the Mannich reaction, and it involves the introduction of a

74 positive charge into the complex tannin structures [13]. During the Mannich reaction, an iminium ion  
75 is produced through the condensation of an aldehyde with an amino compound, which then substitutes  
76 hydrogen in the polyphenolic matrix of tannin [12].

77 The cationization of the anionic polyphenolic structure of tannins allows them to be able to destabilize  
78 anionic contaminants during water purification. Nonetheless, studies indicate that the efficiency of  
79 tannin-based polymers as a coagulant is dictated by several factors including the structure, origin and  
80 chemical modification procedure [12,14]. Grenda et al. [12] examined the influence of the method of  
81 synthesis on the physicochemical properties of tannin coagulants by comparing two cationization  
82 procedures (one step and dual step system) while other parameters and conditions remained constant.  
83 The studies revealed that the dual system recorded a higher reduction in turbidity, colour and COD,  
84 indicating the importance of the synthesis procedure on the overall effectiveness of the tannin  
85 coagulant. Another widely reported factor that determines the performance of these bio-coagulants  
86 during water purification is the effluent characteristics. Effluent properties such as pH, salinity and  
87 organic matter concentration are known to play a significant role in the viability of these coagulants.  
88 For example, the surface charge of colloids and the degree of ionization of polymers are both highly  
89 pH-dependent and, hence, the level of flocculation is considerably affected by the pH of the effluent  
90 [15], whereas salinity is believed to inhibit the chemical activity of polymers by masking their  
91 functional sites and altering their pristine chemical structure [16]. Finally, results from earlier research  
92 have also pointed out the importance of optimum dosage, as a higher coagulant dosage could lead to  
93 a charge reversal, resulting in the restabilization of the suspension [17].

94 This study aimed to investigate the possibility of using Mannich-modified tannin extract from the  
95 bark of spruce (*Picea abies*) as a bio-coagulant in water treatment applications. Although many  
96 scientific papers have been published on the coagulative performance of tannin-based coagulants  
97 from different tree species such as *Acacia mearnsii* (acacia) [12,18,19], *Schinopsis balansae*  
98 (quebracho) [12,18,19] and *Castanea sativa* (castanea) [19], no study was found to have explored the

99 coagulation performance of a bio-coagulant of any spruce species. Spruce bark is a major “side-  
100 stream” waste of many forestry industries, especially in northern European countries. Earlier studies  
101 have reported that the tannin content in spruce bark varies and has been estimated to range between  
102 4% and 15% [20,21]. In this study, two types of spruce bark tannins (the main difference being in the  
103 pulverization method, i.e. spray drying and freeze drying) were chemically modified using two  
104 different amines (diethanolamine and ethanolamine) and compared to the reference material,  
105 quebracho tannin. The study with spruce tannin obtained from the two commonly used pulverization  
106 methods provides new information on the most suitable modification pathway for spruce tannin  
107 extract. The tannins were characterized by X-ray photoelectron spectroscopy (XPS) and ultra-high-  
108 performance liquid chromatography triple quadrupole mass spectrometry to obtain information about  
109 the phenolic compounds present [22,23]. Molecular weights and charge densities were measured for  
110 the obtained bio-coagulants. Previous research studies have rarely reported charge density for  
111 obtained bio-coagulants even though it is one of the most important parameters that determine the  
112 performance of coagulants. Finally, the ability of bio-coagulants to remove turbidity and organic  
113 matter from a kaolin/river water mixture was investigated, and the jar test residue was characterized  
114 by XPS.

115

## 116 **2. Materials and Method**

### 117 **2.1 Materials**

118 Fresh spruce bark for tannin extraction was obtained from Stora Enso Oyj Veitsiluoto mill in Kemi,  
119 Finland. The dry matter content of the bark was approx. 50%. Spray-dried spruce tannin (referred to  
120 hereinafter as SDS) was provided by VTT, the Technical Research Centre of Finland. Spruce bark  
121 was subjected to hot water extraction at 90°C. The obtained tannin extract was then spray-dried at  
122 210-220 mbar with inlet and outlet temperatures of 175 °C and 75 °C respectively, to obtain a tannin  
123 powder with 6 % moisture content [24]. Quebracho tannin (QS-SOL) was kindly supplied by

124 Silvateam (Italy). The chemicals and reagents used for the study were of analytical grade.  
125 Ethanolamine (ETH) and diethanolamine (DEA) were supplied by Sigma Aldrich Chemicals,  
126 Germany. Milli-Q ultrapure water was used throughout this study except where indicated.  
127 Formaldehyde (37% v/w) was produced in France by VWR International while Merck KGaA  
128 Germany manufactured the HCl used for acidification of the synthesized coagulants. The sodium  
129 polyethylenesulphonate (PesNa) and poly-diallyl-dimethyl-ammonium-chloride (PolyDadmac)  
130 solutions used for the measurement of charge density of modified coagulants and total surface charge  
131 of water samples were supplied by BTG instruments AB, Sweden. The NaOH used to adjust the pH  
132 of the jar test water was supplied by VWR Chemicals.

133 Test water for the jar tests was produced by mixing a certain volume of river water and a weighed  
134 amount of kaolin (see section 2.3). The water sample was collected from the Oulu river (Oulu,  
135 Finland). The kaolin used was of sieved fraction size 0.063-0.5 mm and was collected from  
136 Pihlajavaara 64° 48', located in the Kainuu region of Finland [25]. The nutrients and elemental  
137 content of the river water was measured according to standardized methods: Cl, F and SO<sub>4</sub><sup>2-</sup> analysis  
138 was performed by ion chromatography (SFS-EN ISO 10304-1:2009); the NO<sub>2</sub>-N and NO<sub>3</sub>-N in the  
139 river water were analysed using a continuous flow analyser (SFS-EN ISO 13395:1997), the NH<sub>4</sub>-N  
140 and PO<sub>4</sub>-P were also analysed with a continuous flow analyser using the SFS-EN ISO 11732:2005  
141 and SFS-EN ISO 15681-2:2005 methods, respectively. Inductively coupled plasma optical emission  
142 spectrometry (ICP-OES) (SFS-EN ISO 17294-2:2016) was employed to measure the concentration  
143 of the elements.

144

## 145 **2.2 Tannin extraction and pulverization**

146 Tannins were extracted from the fresh spruce bark within 48 h of delivery to avoid degradation and  
147 immobilization of extractable tannin due to prolonged storage. Before tannin extraction, large chips  
148 of wood (pure wood over 5 cm and 0.3 cm in length and thickness, respectively) were removed

149 manually from the spruce bark sample. This was followed by weighing the bark/deionized water of  
150 1:10 ratio (10% w/v) and heated at a temperature of 85 °C, with mechanical stirring at 100 rpm. The  
151 ratio and temperature were selected according to previous research by Kemppainen [24], which  
152 reported a higher percentage of bound sugar in the tannin extract when the temperature was increased  
153 to 90°C and a minimal increase in tannin extract when the bark/water ratio was increased above 1:10  
154 (10% w/v). The mixture was heated for 2 h, after which the tannin extract was decanted. Next, the  
155 extracted tannins and residual bark were effectively separated by centrifuging at 3000 rpm for 10  
156 minutes. Finally, the tannin extract was stored in a plastic capped bottle at -18 °C for further  
157 processing.

158 Before pulverization of the spruce tannins, the tannin concentration in the extract was increased with  
159 the aid of a rotary evaporator. During the tannin concentration process, 300 ml of water was removed  
160 from every 500 ml of the spruce extract in the rotary evaporator at 65 to 70 °C. At the end of the  
161 vacuum evaporation exercise, the tannin concentration of the extract was measured by acid butanol  
162 assay for proanthocyanidins [26], which revealed that the tannin concentration in the extract had  
163 increased from 1442 mg/l to 2650 mg/l. Pulverization of the tannin was performed through freeze  
164 drying with a SCANVAC cool-safe freeze dryer and a completely freeze-dried tannin was obtained  
165 after 16 days (referred to hereinafter as FDS (Freeze-dried spruce)).

166

### 167 **2.3 Kaolin/river water mixture**

168 The kaolin/river water was prepared by spiking 5 l of river water with 1 g of fine kaolin (200 mg/l).  
169 In order to obtain an effective kaolin dispersion, it was important to adjust the pH of the suspension  
170 slightly [13,27]. A stable kaolin suspension was achieved by adjusting the pH of the river water to  
171 7.5 from its initial pH of ~6 with 0.5 M of NaOH before the addition of kaolin. The suspension was  
172 blended by mixing vigorously with a magnetic stirrer at a rate of 800 rpm for 2 h. The characteristics  
173 of the obtained kaolin/river water mixture are shown in Table 3.

## 174 **2.4 Mannich modification of tannin-based coagulant**

175 Modification was performed by adapting the Mannich modification process found in the literature  
176 [12]. The Mannich reaction was initiated by dissolving 2.5 g of powdered tannin in 5 ml of MQ water  
177 at room temperature. The tannin powder was added gradually in batches and stirred with a laboratory  
178 spoon to avoid tannin aggregation and to ensure its complete dissolution. It is important to note that  
179 physical observation of the tannins revealed that the QS-SOL and SDS tannins were composed of  
180 symmetric, fine powdered particles while the freeze-dried tannin was composed of flaky powdered  
181 particles. FDS tannin had the largest volume per gramme but was the most soluble when compared  
182 to the other tannins. The dissolution was preceded by raising the temperature of the tannin solution  
183 to 70 °C before adding 10.7 ml of diethanolamine or 4.9 ml of ethanolamine, and the pH was adjusted  
184 to 6.5 with concentrated HCl (37%) under mechanical agitation by a magnetic stirrer. The temperature  
185 was again raised to 80 °C, and 1.38 ml of formaldehyde was added gradually with a peristaltic pump  
186 over 90 minutes. After the addition of formaldehyde, the temperature of the solution was increased  
187 to 85 °C and mixed continuously for 3 h. At the end of this time, the reaction was stopped by adding  
188 5 ml of MQ water. The modified product was standardized by transferring it to a 50 ml volumetric  
189 flask and filling it up to the mark with MQ water. The solid content of the coagulants obtained before  
190 standardization was ~12% (w/v) and ~14% (w/v) for the DEA and ETH modification respectively  
191 and ~5% (w/v) for both amine modifications after standardization.

192

## 193 **2.5. Characterization of tannin samples, tannin-based coagulants and jar test residue**

194 The non-modified and modified tannin samples were analysed by ultra-performance liquid  
195 chromatography assisted by diode array (UPLC-DAD) and triple quadruple mass spectrometry (MS)  
196 detection. Specifically, we used the group-specific MS/MS analyses of procyanidin and  
197 prodelfphinidin sub-units of proanthocyanidins together with the negative ion full scan analysis in the  
198 ion range of  $m/z$  100-1200, as shown in detail in Engström et al. [22].



199 In addition, tannin samples and jar test residues were studied by XPS. The XPS characterization was  
200 performed with a Thermo Fisher Scientific ESCALAB 250xi using a rotating monochromatic Al  
201 anode generating an X-ray beam at 1486.6 eV. The XPS data were analysed with Avantage software,  
202 and the charge correction was performed by setting the binding energy (BE) of adventitious carbon  
203 to 284.8 eV while the Shirley function was used to subtract the background. The charge density  
204 measurement of the coagulants was performed with a Mütek particle charge detector PCD 03 pH  
205 (Hersching, Germany) by titrating with PesNa. The molecular weight (MW) of the coagulants was  
206 determined through high-pressure size exclusion chromatography (HP-SEC, Aligent 1100 Series  
207 Liquid Chromatography, Hewlett-Packard, USA) equipped with a Shimadzu RID-10A detector. 0.1  
208 M of NaNO<sub>3</sub> was used as eluent with a flow rate of 0.5 ml/min in Ultrahydrogel (500 + 125) columns.  
209

## 210 **2.6. Water analyses**

211 The turbidity of the water samples was measured with a HACH 2100Q portable turbidimeter while  
212 the UV absorbance of the water samples was measured with a Shimadzu ultraviolet spectrometer at  
213 254 nm. The total surface charge of the water samples was determined with a Mütek particle charge  
214 detector by titrating 10 ml of the supernatants with PesNa or Poly-Dadmac titrant. The pH of the  
215 solutions was determined with a VWR phenomenal pH 1000 L. For river water and kaolin/river water,  
216 the SUVA value (specific ultraviolet absorbance) was calculated based on the measurement of the  
217 ultraviolet absorbance (UVA) of the water at an optical wavelength of 254 nm and the concentrations  
218 of dissolved organic carbon (DOC) according to EPA methods [28]. DOC measurement was carried  
219 out with a SIEVERS 900 portable TOC analyser. Before the UVA and DOC measurements, the  
220 samples were filtered through a 0.45 µm filter membrane (VWR, polyethersulphone membrane).

221

222

223

## 224 **2.7. Coagulation experiments**

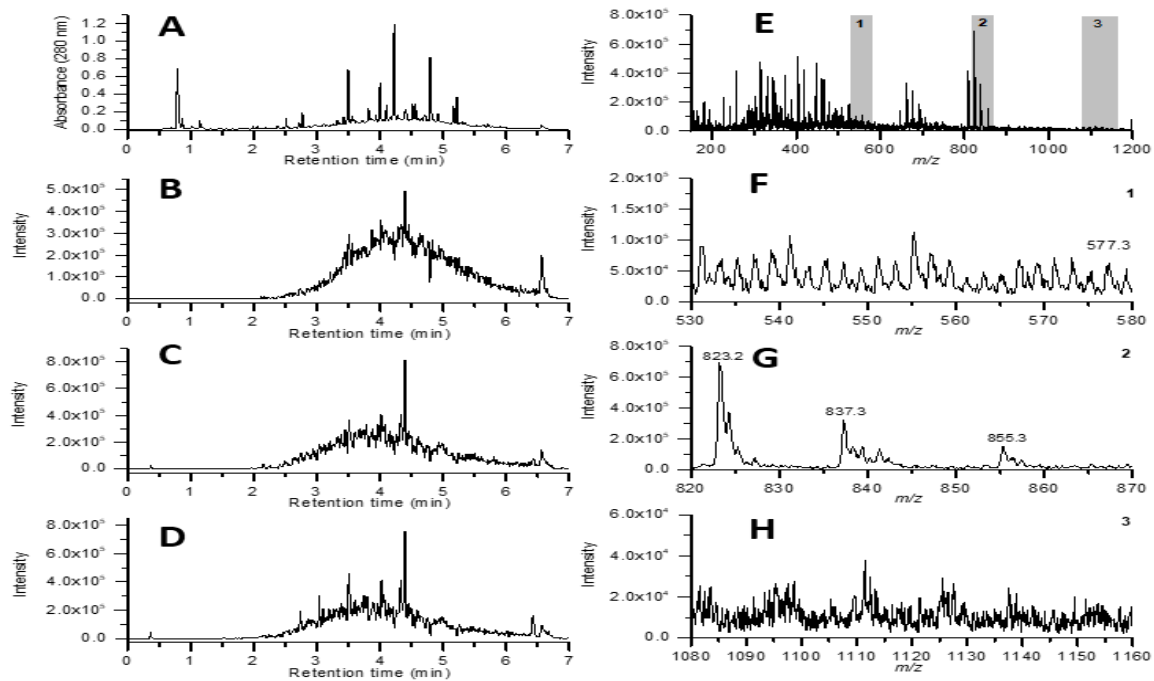
225 For the coagulation experiments, Kemira flocculator 2000 jar test apparatus with conventional 1.0 l  
226 glass beakers was used. During the tests, the beakers were filled with the kaolin/river water to 800  
227 ml. The suspension was then dosed with a different tannin coagulant and the solutions were subjected  
228 to rapid mixing at 150 rpm for 1 minute, slow mixing at 40 rpm for 20 minutes and sedimentation for  
229 30 minutes. Turbidity samples were taken from 5 cm below the surface after the sedimentation period  
230 and were measured immediately. This was then followed by the extraction of 200 ml of the  
231 supernatant from a point 3 cm below the surface of the test water sample for other water analyses. Jar  
232 test residues were obtained by carefully decanting the supernatant from the sedimented flocs in the  
233 glass beakers. The recovered flocs were centrifuged at 2500 rpm for 10 minutes and the supernatants  
234 were again decanted. The flocs were oven-dried at 40 °C for 8 hours to obtain dried residues.

235

## 236 **3. Results and Discussions**

### 237 **3.1. Characterization of tannin extracts**

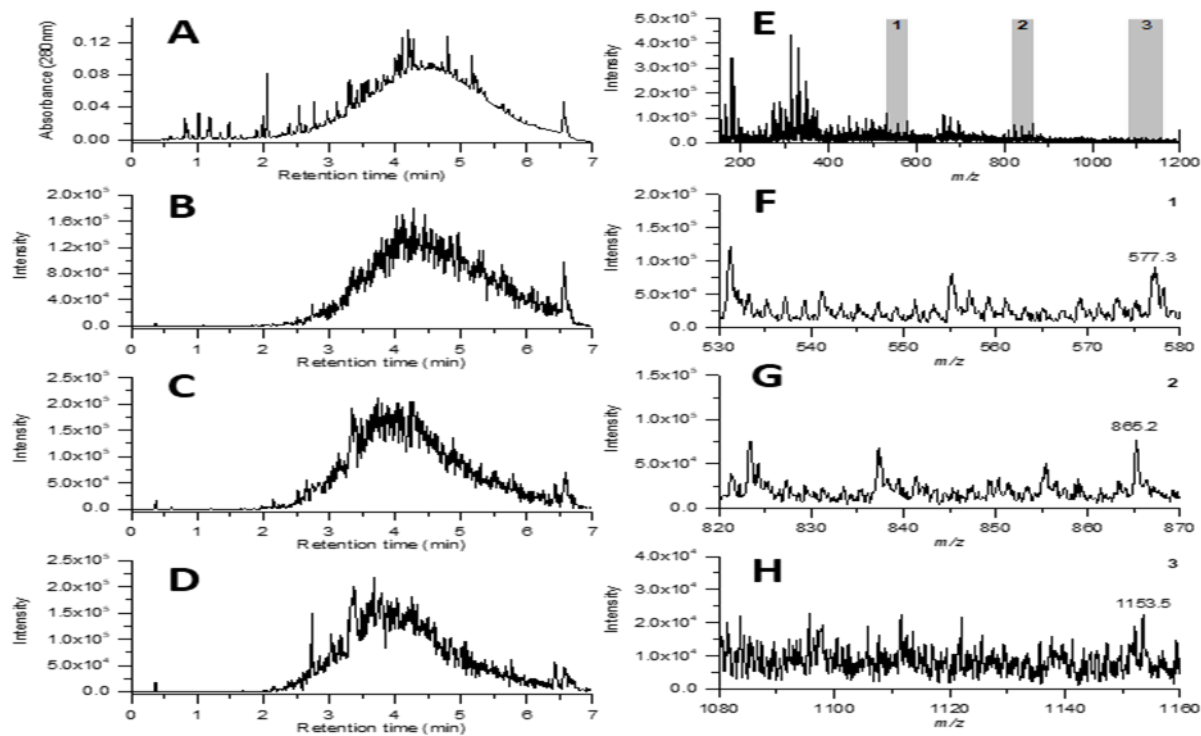
238 The UPLC-DAD-MS/MS analyses did indeed reveal that the polyphenol composition of the FDS and  
239 SDS spruce samples and that of QS-SOL tannin were quite different. The UPLC chromatograms  
240 show that the FDS sample contained numerous simple phenolic compounds, represented by sharp  
241 chromatographic peaks on top of the chromatographic hump of proanthocyanidins (PA) (Fig. 1A).  
242 On the other hand, the SDS sample was dominated by the proanthocyanidin hump and had only low  
243 levels of simple phenolics (Fig. 2A). The QS-SOL tannin sample contained the proanthocyanidin  
244 hump, together with some simple phenolics (Fig. 3A) such as gallic acid at 1.2 mins.



245

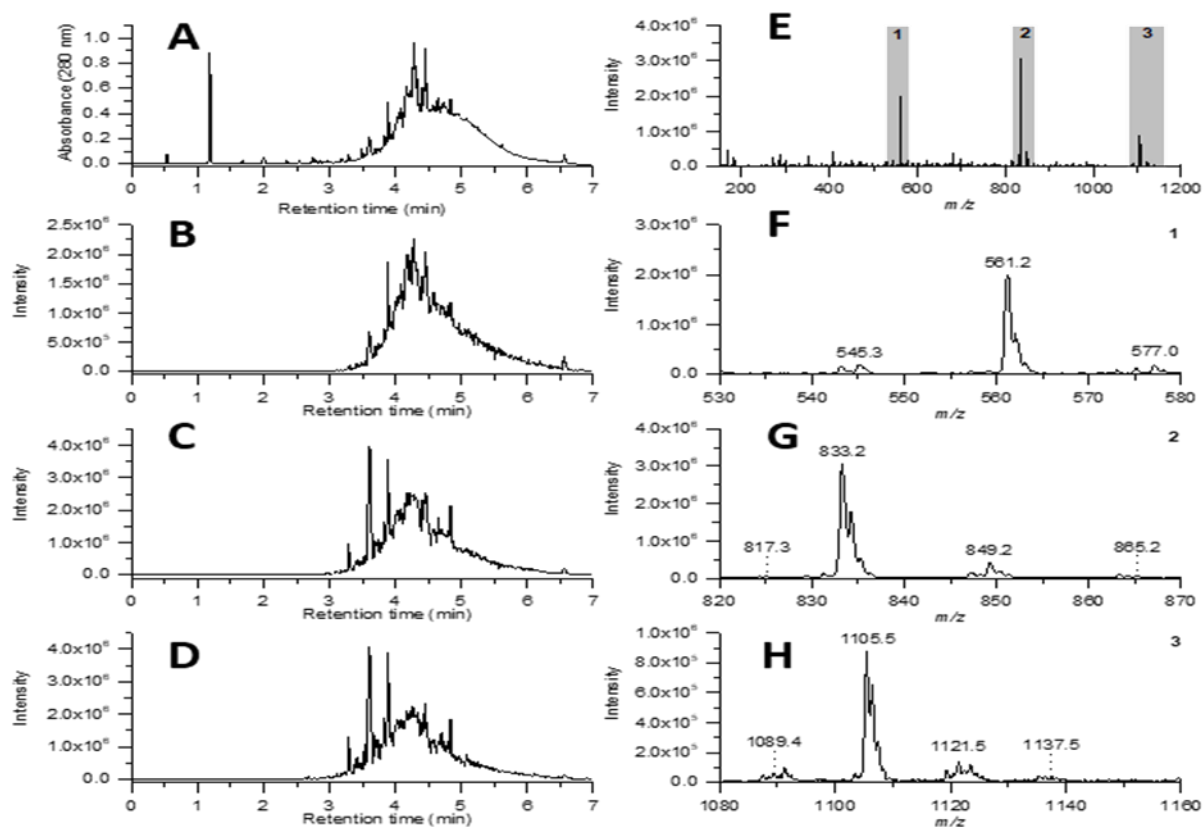
246 **Figure 1.** The UPLC-DAD-MS characteristics of the FDS sample: (A) the UPLC chromatogram at  
 247 280 nm, (B-D) procyanidin MS/MS fingerprints at different cone voltages, (E) full scan spectra, (F-  
 248 H) zoomed mass spectra for the m/z regions for dimeric, trimeric and tetrameric procyanidins.

249



250

251 **Figure 2.** The UPLC-DAD-MS characteristics of the SDS sample: (A) the UPLC chromatogram at  
 252 280 nm, (B-D) procyanidin MS/MS fingerprints at different cone voltages, (E) full scan spectra, (F-  
 253 H) zoomed mass spectra for the  $m/z$  regions for dimeric, trimeric and tetrameric procyanidins.  
 254



255  
 256 **Figure 3.** The UPLC-DAD-MS characteristics of the QS-SOL sample: (A) the UPLC chromatogram  
 257 at 280 nm, (B-D) procyanidin MS/MS fingerprints at different cone voltages, (E) full scan spectra,  
 258 (F-H) zoomed mass spectra for the  $m/z$  regions for dimeric, trimeric and tetrameric procyanidins.  
 259

260 The MS/MS method produced clear procyanidin fingerprints for all three samples (Figs. 1B-D, 2B-  
 261 D, 3B-D), while the prodelphinidin fingerprints were much smaller in intensity (data not shown). The  
 262 more numerous ions in the full scan ESI-MS spectra (panels E in Figs. 1-3) of the FDS tannin sample  
 263 further confirmed that it contained simpler phenolics than the SDS tannin. The ion regions  
 264 corresponding to the procyanidin dimers ( $m/z$  577), trimers ( $m/z$  865) and tetramers ( $m/z$  1153) are

265 shown in panel E and zoomed in panels F-H, highlighting the fact that only the SDS tannin sample  
 266 produced clear signals for these PA oligomers. For the QS-SOL tannin, these ions were scarce,  
 267 whereas the ions at  $m/z$  561, 833 and 1105 dominated the full scan mass spectra. These correspond  
 268 to 5-deoxy PA dimers, trimers and tetramers showing that the quebracho proanthocyanidins differ a  
 269 lot from the spruce proanthocyanidins. These data allowed us to calculate the PA concentration, the  
 270 mean degree of polymerization (mDP) and the PC/PD ratio for the spruce (Table 1), but not for the  
 271 quebracho PAs, since the Engström method [22] is not suited for detecting 5-deoxy PAs.  
 272 Nevertheless, the QS-SOL tannin sample clearly had the highest PA content, although not all of its  
 273 PA sub-units were included in the quantification.

274

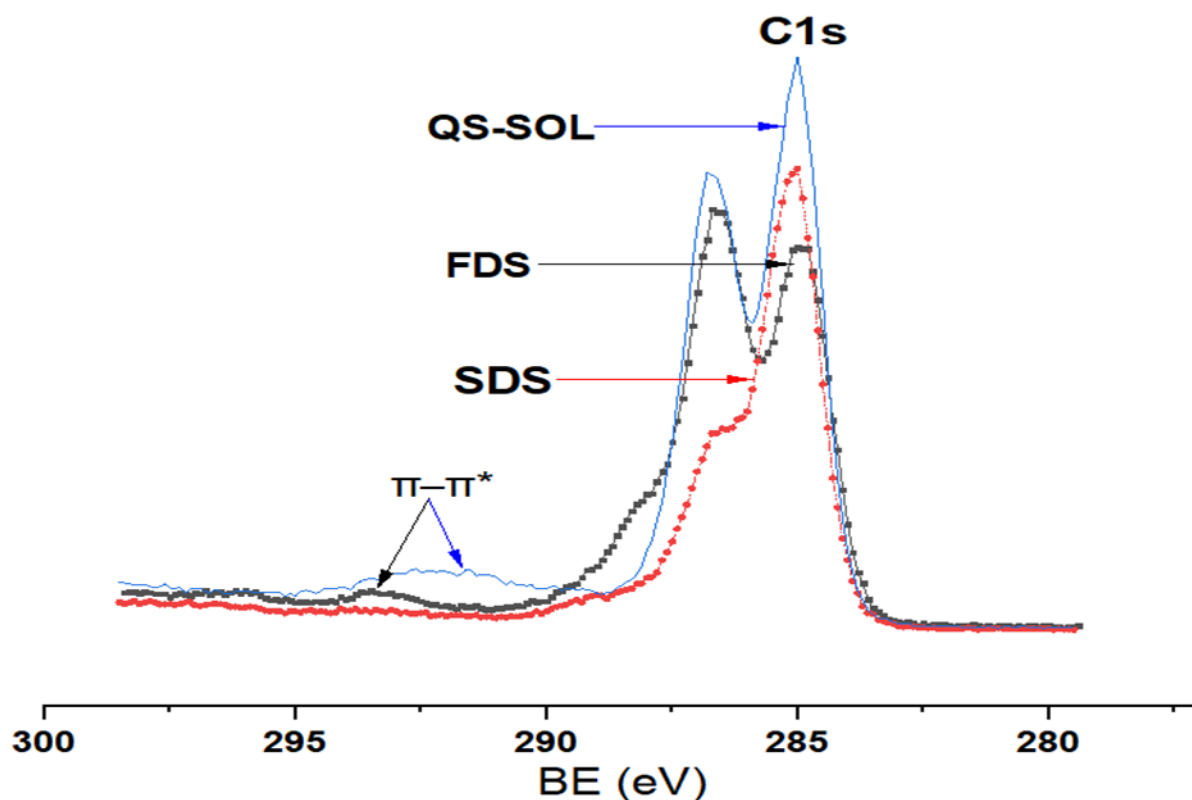
275 **Table 1.** The proanthocyanidin (PA) content, the procyanidin (PC) content and the mean degree of  
 276 polymerization (mDP) of the proanthocyanidins in the FDS, SDS and QS-SOL samples.

Sample	PA (mg/g)	PC-%	mDP
FDS	17.3	95	4.4
SDS	11.7	81	2.8
QS-SOL	64.6	-	-

277

278 The XPS wide spectra for the tannin samples (Supplementary Fig. S1) indicate that carbon (C1s 284.8  
 279 eV) and oxygen (O1s 532 eV) were the only detectable elements. Although the presence of a  
 280 significant amount of hydrogen in the tannin matrix is generally known, XPS is unable to detect  
 281 hydrogen. Moreover, with that in mind, it can be said that the XPS spectra are in confirmatory  
 282 agreement with the generally represented chemical composition (C, H and O) of tannin. The C1s  
 283 spectrum for the tannin samples presented in Fig. 4 clearly revealed two main peaks at 284.8 eV and  
 284 at 286.5 eV, which can be assigned to C–C, C=C and C–H (C1 component) and C–OH and C–O (C2  
 285 component), respectively [29]. A minor peak was observed at ~288.0 eV with the spruce tannin

286 samples (Fig. 4, Supplementary Fig. S2A and 2B) and can be assigned to C=O (C3 component) [30].  
287 The presence of the C3 carbon component was attributed to the presence of non-stilbene glycosidic  
288 and polymeric sugars in spruce tannin extracts [31].  
289 Furthermore, a shake-up satellite, namely the  $\pi$ - $\pi^*$  component, was observed in the spectra of the  
290 FDS and QS-SOL tannins at BE  $\sim$ 293 eV and  $\sim$ 292 eV, respectively, but was featureless in the SDS  
291 tannin. The  $\pi$  bond is a characteristic feature of the  $sp^2$  aromatic carbon constituent, and its intensity  
292 is dictated by the contribution of C=C bonding to the overall C1s spectrum [32]. However, it is  
293 essential to note that the shake-up satellite was more pronounced in the QS-SOL, which reflects a  
294 higher amount of  $sp^2$  aromatic carbon.



295  
296 **Figure 4.** Overlaying comparison of C1s spectra for tannin samples.

297  
298 The percentage amount of the C2 component (C-OH and C-O) was 42.2% for FDS, 43.9% for QS-  
299 SOL and 27.5% for SDS tannins (Supplementary Fig. S2). For QS-SOL tannin, the higher amount of  
300 phenolic groups could have contributed to the prominence of the C2 component, which is also in

301 agreement with the proanthocyanidin content (Table 1). Kemppainen [24] had earlier reported that  
302 quebracho tannin contained almost seven times more phenolic groups than spruce tannin. The higher  
303 amount of the C2 component in the FDS tannin sample as compared to SDS tannin could be attributed  
304 to the greater amount of proanthocyanidins (Table 1) and the presence of simple phenolics (Fig. 1).

305

### 306 **3.2. Characterization of tannin coagulants**

307 The modified bio-coagulants did not contain detectable levels of the original PAs, but their UPLC-  
308 chromatograms showed a retention time shift for the “tannin hump” (Fig. S3 and S4). The MS/MS  
309 fingerprints did not show any traces of PC or PD building blocks (data not shown), neither did the  
310 full scan mass spectra witness any ions specific to the natural PAs found in native tannin samples.  
311 This suggested that the modification reactions were successful.

312 The average molecular weights of all the coagulants were low, at approximately 390 g/mol (Fig. S5,  
313 ~37.5 min). However, a similar value has been reported for a commercial tannin coagulant [33]. Table  
314 2 shows the charge density of the modified coagulants with two different amine sources. The  
315 quebracho tannin coagulant had the highest charge density with both amine sources. It could be  
316 deduced that the charge densities of the ethanolamine-modified coagulants were higher than their  
317 diethanolamine counterparts. Comparison of charge densities with other studies is difficult since most  
318 of the studies do not report these values for tannin-based coagulants and different aminomethylation  
319 pathways were commonly used [12,18,19]. However, at least Gang [13] reported a similar charge  
320 density for a tannin-based coagulant (3.1 meq/g) prepared through the Mannich modification of  
321 tannin (from an unknown source) with ammonium chloride and formaldehyde.

322

323

324

325 **Table 2.** Charge densities of the tannin coagulants. Range of charge densities represents the deviation  
326 in two repeats.

<b>Bio-coagulant</b>	<b>Charge density (meq/g)</b>
FDS-DEA	$0.83 \pm 0.10$
SDS-DEA	$0.85 \pm 0.04$
QS-SOL-DEA	$2.79 \pm 0.30$
FDS-ETH	$1.65 \pm 0.31$
SDS-ETH	$1.88 \pm 0.17$
QS-SOL-ETH	$3.84 \pm 0.41$

327

328 According to the charge density results in Table 2, the FDS was discovered to have the lowest charge  
329 density after both amine modifications. Aside from the variation in pulverization methods, one  
330 recorded difference between the Mannich-modified spruce tannins is that the FDS tannin came from  
331 a tree harvested towards the end of the winter season, while the SDS tannin was extracted from  
332 summer-harvested trees. Earlier researchers [24,34] have reported a higher amount of glucose units  
333 in the tannins of winter-felled spruce trees while a higher tannin content was recorded in the bark of  
334 summer-harvested trees.

335

### 336 **3.3 Coagulative performance of tannin coagulants**

337 Although the lower efficiency of commercial tannin coagulants in removing humics as compared to  
338 iron coagulants has been reported earlier [33], river water was selected for the test water, but only  
339 after spiking with kaolin. The characteristics of the river water sample before and after spiking with  
340 kaolin are shown in Table 3. Results from these analyses indicate that the river water had a low initial  
341 particle and colloid amount (low turbidity) but the turbidity was increased by spiking with fine kaolin.  
342 The high  $UV_{254}$  and SUVA values of the river water indicate the presence of high aromatic and humic



343 type organic substances. The analysed parameters also showed that spiking with kaolin had a minimal  
 344 effect on the pristine characteristics of the river water.

345

346 **Table 3.** Characteristics of river water and kaolin/river water.

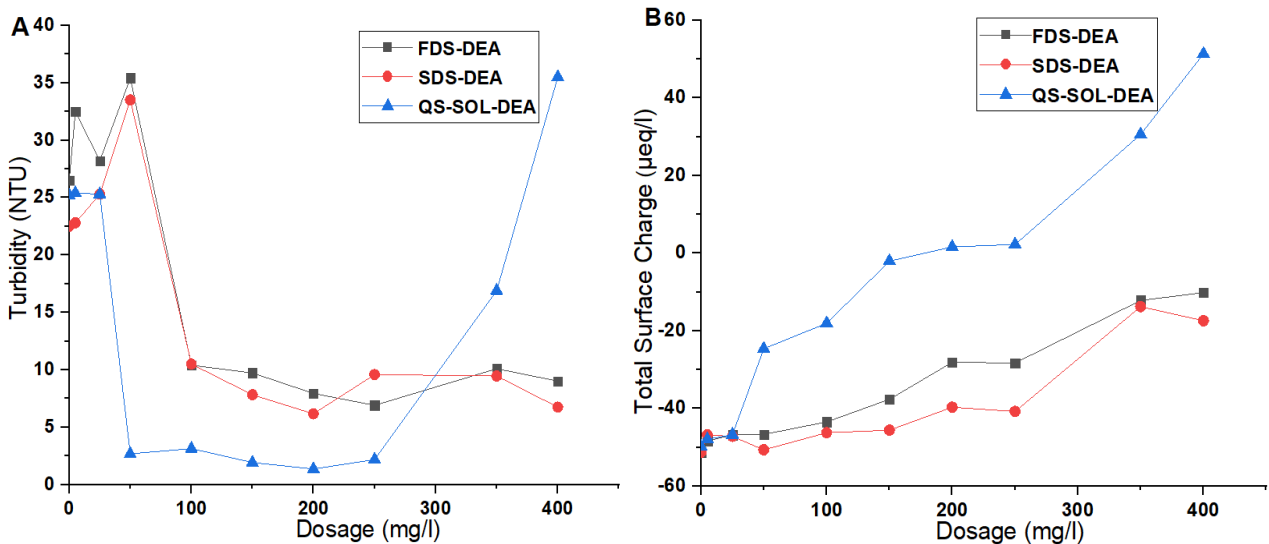
	<b>Turbidity</b>	<b>pH</b>	<b>UV<sub>254</sub></b>	<b>DOC</b>	<b>TSC</b>	<b>SUVA</b>
	<b>(NTU)</b>		<b>1/cm</b>	<b>(mg/l)</b>	<b>(µeq/l)</b>	<b>(l/mg-M)</b>
<b>River water</b>	1.3	6.1	0.328	11.7	-45.41	2.803
<b>Kaolin/river water</b>	25.2	7.5	0.363	11.2	-50.65	3.135

347

348 The Mannich-modified tannins were able to reduce the turbidity of the kaolin/river water well. Fig.  
 349 5A and Fig. 5C show the relationship between various coagulant dosages and turbidity removal. A  
 350 detailed comparison of their performance showed that, during the clarification test, the quebracho  
 351 tannin coagulant slightly outperformed the spruce tannin coagulants in both amine modifications by  
 352 achieving >90% turbidity reduction at a dosage of 50 mg/l for the DEA modification (Fig. 5A) and  
 353 25 mg/l for the ETH modification (Fig. 5C). However, these results also indicate that the modified  
 354 spruce tannin coagulants recorded an impressive turbidity reduction at 100 mg/l dosages for the DEA  
 355 modifications and 50 mg/l for the ETH modifications. This trend was observed to be in contrast with  
 356 earlier publications where it was claimed that DEA Mannich-modified quebracho tannin exhibited  
 357 better performance over the ETH modification for removal of dyes and surfactants from synthetic  
 358 effluents [18,19]. However, the performance of bio-based coagulants has been reported to vary in  
 359 different effluents due to the distinctive characteristics of the effluent concerned [15,16]. In addition,  
 360 the turbidity values was noticed to be stable over a wide range of dosage with spruce tannin coagulants  
 361 in both amine modifications. This could be of immense benefit when used in applications where the  
 362 coagulant dose is constant, and dosing is not based on fluctuations in water quality.

363 Fig 5B and D show the total surface charge of water samples after the jar test for tannin coagulants  
 364 using DEA and ETH as amines, respectively. The results indicate that all the water samples tested  
 365 with both classes of coagulant modification exhibited a decrease in surface charge (i.e. the TSC  
 366 became less negative) as the coagulant dosage increased. When ETH was used as an amine, a lower  
 367 dose of coagulant was required to achieve an optimal dosage in comparison with DEA. The results  
 368 for the total surface charge of the water samples after jar testing also revealed that, when overdosed  
 369 with a higher amount of coagulant, the use of QS-SOL tannin coagulants resulted in a high cationic  
 370 charge whereas the water treated with spruce tannin coagulants had a TSC of close to zero. The trends  
 371 portray a strong interrelationship between the charge density (Table 3), total surface charge and  
 372 turbidity removal. Thus, it is reasonable to conclude that the higher the charge density of coagulant,  
 373 the lower the dose required to neutralize negatively charged colloids. This trend is supported by Gang  
 374 [13], who also tested Mannich-modified tannin (unknown source) in kaolin-spiked river water and  
 375 concluded that the optimal dosage of tannin-based coagulants was highly dependent on its charge  
 376 density for effluents with a charge demand.

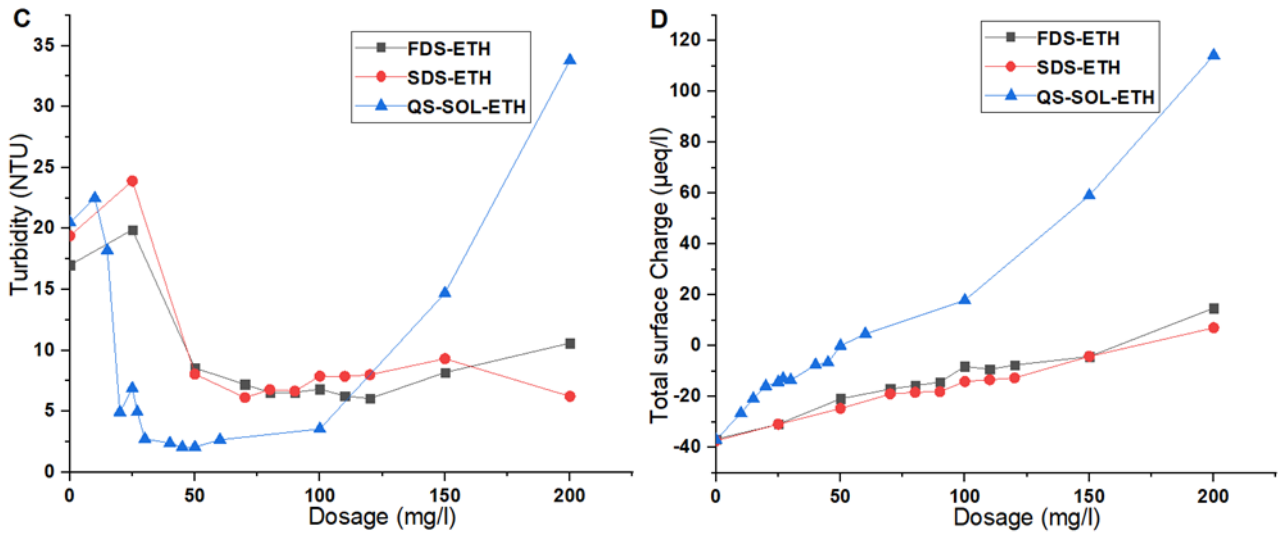
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382 **Figure 5.** (A) Turbidity clarification performances of bio-coagulants with DEA as amine source, (B)  
 383 total surface charge of water sample with DEA as amine source, (C) Turbidity clarification  
 384 performance of bio-coagulants with ETH as amine source, (D) total surface charge of water sample  
 385 with ETH as amine source.

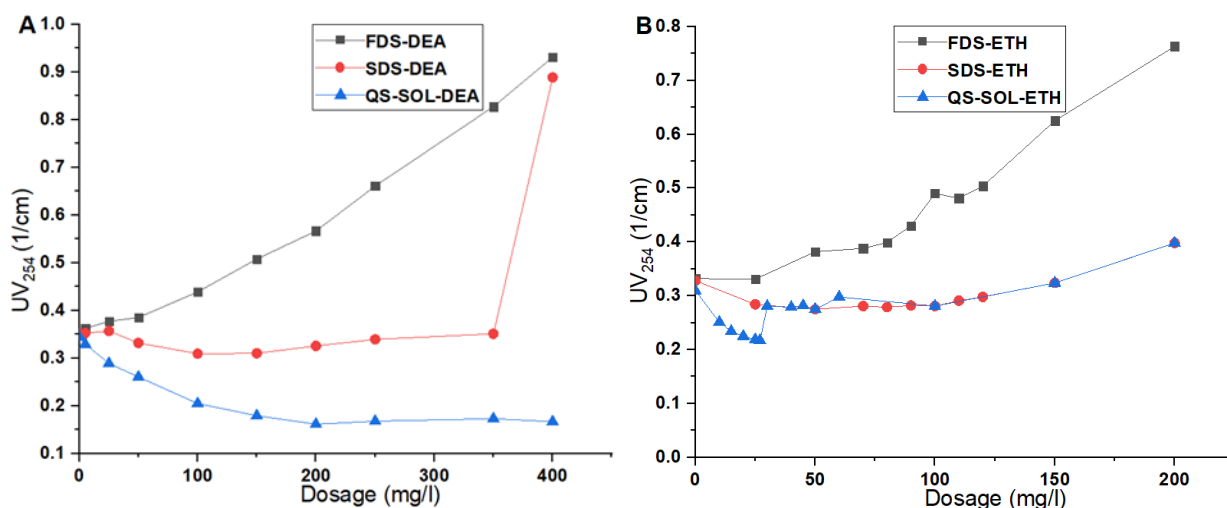
386

387 Figs. 6A and B plot the  $UV_{254}$  as a function of the dosage of DEA and ETH modified coagulants,  
 388 respectively. For DEA modification, the results showed that  $UV_{254}$  absorbance decreased with QS-  
 389 SOL and SDS as the coagulant dosage increased, but the absorbance value of SDS-DEA-treated  
 390 samples was observed to increase after exceeding a dosage of 100 mg/l. In contrast, the  $UV_{254}$   
 391 absorbance value of the FDS-DEA treated water samples increased as the dosage increased, which  
 392 indicates an increase in organic compounds [35,36]. The results of the  $UV_{254}$  absorbance trends  
 393 obtained from the ETH tannin modifications were similar to those of the DEA modifications. The  
 394 only exception was that the  $UV_{254}$  of the QS-SOL-ETH-treated water samples increased after reaching  
 395 the optimal dosage of 25 mg/l, which signifies that optimal dosage is a combination of turbidity and  
 396  $UV_{254}$ .

397

398

399



400

401 **Figure 6.** UV absorbance of the treated water with (A) DEA as amine source, (B) with ETH as amine  
 402 source.

403

404 In general, the results revealed that quebracho tannin coagulants possess better coagulative potential  
 405 in both amine modifications in comparison with the spruce tannin coagulants. The variation in  
 406 performance could be best attributed to two factors. Since the primary mechanism behind the  
 407 aminomethylation process is the addition of an iminium ion to the phenolic ring of tannin, the  
 408 availability of more phenolic groups in the quebracho tannin provides more active sites for the  
 409 iminium ion to substitute hydrogen in the aromatic tannin structure. Another factor that might have  
 410 contributed to the poorer performance of the spruce tannin coagulants is the presence of  
 411 impurities. Tannin extracts from spruce bark mostly consist of ash and carbohydrates [24,37] and  
 412 lower levels of tannins, which leads to the formation of undesired by-products during  
 413 aminomethylation [19]. Zhang and Gellerstedt [37] showed that one in two of the flavonoid  
 414 monomers in the condensed tannin of spruce was covalently bonded to glucose; here we showed that  
 415 FDS spruce tannin contains more flavonoid glycosides than SDS tannin (peaks at 4-6 mins shown in  
 416 Figs. 1-2). Kemppainen [24] also corroborated this claim but further observed that the portion of  
 417 monosaccharides such as glucose in spruce tannin extract was higher in winter-felled trees than in  
 418 those harvested during the summer season. These by-products and impurities with different chemical

419 structures and properties have a high probability of interfering with the Mannich reaction, and it  
420 appears that the aminomethylation might not have proceeded so efficiently with the spruce tannin  
421 extract and consequently, reducing the regioselectivity of the Mannich modification [19]. However,  
422 purification of the extract from impurities could be performed to overcome this drawback. Covalently  
423 bonded carbohydrates such as polysaccharides and disaccharides can be hydrolyzed to  
424 monosaccharides with hydrolytic enzymes, which could be dissociated from tannin through size-  
425 based separation techniques such as ultrafiltration [24]. It should be stated that purification of the  
426 extract would result in high cost, and thus should be avoided. Nevertheless, this study has proved that  
427 spruce-based tannin coagulants also performed efficiently, and further optimization of both extraction  
428 and Mannich modification might improve their properties.

429

### 430 **3.4 Characterization of jar test residues**

431 XPS analysis of the jar test residues revealed some changes induced by the Mannich reaction on the  
432 pristine structure of the tannins. The XPS spectra for the residues confirmed the presence of Al, Si  
433 and N as additional elements to the tannin structure for all bio-coagulants. Al and Si derived mainly  
434 from the kaolin while N originated from the amine used in the Mannich reaction. Furthermore, the  
435 spectra for the residues revealed the presence of barium in some samples. In order to ascertain the  
436 origin of the barium, a detailed characterization of nutrient and elemental content was performed on  
437 the river water. The results from the analysis (Supplementary Table S1) showed that Ba was present  
438 in low concentration (7.2  $\mu\text{g/l}$ ), which indicates that the bio-coagulants might possess a high affinity  
439 for Ba. The reason behind this trend was not investigated in this study.

440 The N1s spectra for the spruce tannin coagulant residues in the DEA modifications had two visible  
441 peak components, intense peaks at  $\sim 399.9$  eV and less intense peaks at  $\sim 401.9$  eV, while that of  
442 quebracho is the reverse (Supplementary Fig. S6 (A-C)). The peak at binding energy  $\sim 399.9$  eV  
443 indicates the presence of non-protonated amine (N1 component) and the second peak at  $\sim 401.9$  eV

444 is assigned to protonated amine (N2 component) [38]. On the other hand, the N1s spectra of the jar  
445 test residues of the ETH modifications for spruce tannins were observed to have a stronger peak at  
446 ~399.9 eV, assigned to non-protonated amine. However, peak fitting (Supplementary Fig. S6(D-F))  
447 clearly revealed the presence of protonated amine. Fitting of the C1s peaks of the coagulant residues  
448 showed the presence of three components at ~284.8 eV (C–C, C–H, C=C), ~286 eV (C–O, C–N) and  
449 ~288 eV (O–C–O, C=O) (Supplementary Fig. S7) [28,30]. A visual comparison of the C1s spectra of  
450 the tannin samples (Fig. S2) and jar test residues (Fig. S7) shows a decrease in the proportion of the  
451 component at lower BE (C–C, C–H, C=C), consistent with the disappearance of the  $\pi$ - $\pi^*$  shake-up  
452 peak in the tannin samples associated with the aromatic conjugation [32,39].

453

#### 454 **4. Conclusions**

455 The existence of companies producing tannin coagulants and many industrial applications of tannin  
456 coagulants are clear indicators that tannin-based coagulants are a viable option for real application in  
457 water treatment. This present study has demonstrated that tannins extracted from spruce bark can be  
458 cationized to produce efficient bio-based coagulants. For application in surface water with a positive  
459 charge demand, ethanolamine seems to be a better amine source to combine with formaldehyde in  
460 the Mannich reaction. Bio-coagulants made from spruce tannins, in comparison with  
461 quebracho, showed some lower charge densities and lower turbidity removal, which were attributed  
462 to two factors: a reduction in the regioselectivity of the Mannich reaction for spruce tannins due to  
463 the presence of impurities and a smaller amount of phenolic groups to accept the iminium ions  
464 generated during the Mannich reaction. Nonetheless, the spruce tannin coagulants could still be  
465 considered a good product because it possessed more stable residual turbidity and a TSC close to zero  
466 over a wide range of dosage; more importantly, it could be further optimized to improve its  
467 applicability as a bio-based coagulant since only two amines were tested during this study.

468 From the two spruce tannin pulverization methods considered in this study, the water treatment results  
469 obtained indicate that spray drying is the more suitable method. Freeze drying is a much slower  
470 process and a prolonged pulverization process might have negatively affected the quality of the  
471 tannin. Furthermore, variation in the season of tree harvesting was also considered to have affected  
472 the performance of the freeze-dried tannin, as the bark of trees harvested in the winter contains a  
473 smaller amount of tannin and a greater amount of glucose. However, the PA amount was higher in  
474 the freeze-dried tannin sample and the freeze-dried tannin also had higher solubility in the  
475 modification. Thus, more studies are required to optimize the extraction and modification process of  
476 spruce tannins.

477

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482

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